



Research Papers

Feeding–Danger Trade-Offs Underlie Stopover Site Selection by Migrants

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ABSTRACT. To migrate successfully, birds need to store adequate fat reserves to fuel each leg of the journey. Migrants acquire their fuel reserves at stopover sites; this often entails exposure to predators. Therefore, the safety attributes of sites may be as important as the feeding opportunities. Furthermore, site choice might depend on fuel load, with lean birds more willing to accept danger to obtain good feeding. Here, we evaluate the factors underlying stopover-site usage by migrant Western Sandpipers (*Calidris mauri*) on a landscape scale. We measured the food and danger attributes of 17 potential stopover sites in the Strait of Georgia and Puget Sound region. We used logistic regression models to test whether food, safety, or both were best able to predict usage of these sites by Western Sandpipers. Eight of the 17 sites were used by sandpipers on migration. Generally, sites that were high in food and safety were used, whereas sites that were low in food and safety were not. However, dangerous sites were used if there was ample food abundance, and sites with low food abundance were used if they were safe. The model including both food and safety best-predicted site usage by sandpipers. Furthermore, lean sandpipers used the most dangerous sites, whereas heavier birds (which do not need to risk feeding in dangerous locations) used safer sites. This study demonstrates that both food and danger attributes are considered by migrant birds when selecting stopover sites, thus both these attributes should be considered to prioritize and manage stopover sites for conservation.

Key Words: *Calidris mauri*; food abundance; migration; predation danger; stopover site conservation; trade-off hypothesis; Western Sandpiper

INTRODUCTION

One challenge to understanding the causes underlying population declines of Neotropical migrants is that declines may be caused by a suite of factors operating at various stages of a migrant's life cycle. Most research on factors limiting these populations (i.e., reproduction and mortality) has focused on breeding and non-breeding stages (Robinson et al. 1995, Burke and Nol 1998, Zanette et al. 2000, Rodewald and Yahner 2001, Gill et al. 2001, Latta and Faaborg 2002, Whitfield 2003a, Stillman et al. 2005b), but comparatively less work has been carried out on limiting factors during migration (Hutto 1985, Moore et al. 1995) even

though migration appears to be a time of high mortality (Sillett and Holmes 2002, Baker et al. 2004). In light of this, biologists have advocated that the protection of high-quality stopover habitat is essential to conserve migratory bird species (Myers et al. 1987, Skagen and Knopf 1993, Donovan et al. 2002, Mehlman et al. 2005).

Conserving stopover sites requires understanding what makes a site important. Mehlman et al. (2005) propose a conceptual framework to prioritize stopover sites based on their capacity to meet migrants' needs at a given point in space and time and facilitate their successful migration. The authors suggest that criteria to classify and prioritize

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stopover sites for conservation should include: 1) ecological context (e.g., proximity to ecological barriers, degree of spatial isolation, etc.); 2) intrinsic characteristics (e.g., food abundance, safety from predators); and 3) migrant use (e.g., relative abundance, including frequency and consistency of use as a stopover site). Simply put, to identify appropriate stopover sites for conservation, an understanding of the factors underlying their usage on a broad scale is required.

To migrate successfully, sufficient energy reserves are needed to fuel long-distance flights. In light of the rapid, impressive fuelling by long-distance migrants, high-quality stopover sites are generally considered as those that can provide the required energy. Therefore, much research on stopover-site usage has focused on the abundance and distribution of food (Schneider and Harrington 1981, Hicklin and Smith 1984, Loria and Moore 1990, Colwell and Landrum 1993, Piersma et al. 1994, Baker et al. 2004, Placyk and Harrington 2004).

More recently, investigators have begun to recognize a role for predation danger in shaping where, when, and how quickly migrants travel, and how they select and use stopover sites (Moore 1994, Weber et al. 1998, Clark and Butler 1999, Lank et al. 2003). For example, migrants may adjust their fat loads (Ydenberg et al. 2002, 2004, Schmaljohann and Dierschke 2005) and alter the amount of time spent being vigilant (Cimprich et al. 2005, Pomeroy 2006) with changing levels of predation danger, or may avoid feeding in dangerous places (Sapir et al. 2004, Cimprich et al. 2005, Pomeroy 2006, Pomeroy et al. 2006). Feeding–danger trade-offs are central to decisions made by foragers; in fact, this fundamental trade-off can affect processes ranging from individual fitness to population- and community-level processes (i.e., Cresswell 2008, and references therein). At stopover sites, greater danger makes foragers behave more cautiously, whereas higher food availability makes foragers more accepting of danger.

Although these studies inform us of the attributes that influence within-site habitat selection by migrants, they do not necessarily convey which attributes migrants use to select a site at which to stopover (Hutto 1985). Migrants likely use site-level attributes to determine if the site is suitable, and later adjust within-site usage to optimize fuel deposition. Large-scale studies are needed to understand how food and danger affect selection

and usage of stopover sites by migrants. Furthermore, the physical attributes or current state (fuel load) of individuals might influence their stopover-site selection. For example, lean migrants may be more willing to risk using dangerous sites that offer greater feeding benefits, whereas birds with ample fat reserves might use sites that offer greater safety (e.g., Ydenberg et al. 2002).

Here we designed a landscape-scale field study to assess the food and danger attributes of a variety of potential stopover sites to evaluate the factors underlying their usage by Western Sandpipers (*Calidris mauri*). We included stopover sites known to be used by Western Sandpipers, as well as sites that appeared suitable but were not used (criteria to determine “used” and “unused” sites are given below). We predicted that if usage of stopover sites depends on either food abundance or safety from predators alone, then sandpipers will use sites with high food abundance or safety and avoid sites with low food abundance or safety. If usage of stopover sites by sandpipers reflects a trade-off between food abundance and predation danger, then we predict that migrant sandpipers will use stopover sites high in both food and safety and not use sites that are low in both food and safety. Furthermore, birds should use dangerous sites only if they offer high food abundance and sites low in food abundance only if they are safe from predators. If the individual state (fuel load) of migrant sandpipers affects stopover-site usage, then we predict that lean birds should be willing to use dangerous sites whereas heavy birds with ample fuel reserves will use only sites that offer safety from predators. Understanding how these factors affect stopover-site usage allows us to classify the importance of sites to prioritize for conservation (cf. Mehlman et al. 2005).

METHODS

Study System

Western Sandpipers are small (25–30 g on migration) calidrine shorebirds that migrate along the Pacific Coast from non-breeding grounds between northern Oregon and Peru, to breeding grounds in western and northern Alaska and eastern Siberia (Wilson 1994). En route, Western Sandpipers stopover on mud- and sandflats in the Georgia Basin (British Columbia, Canada) and Puget Sound (Washington, USA) regions (Butler et al. 1987, Buchanan 1988, Iverson et al. 1996,

Evenson and Buchanan 1997) during mid-April to mid-May on northward migration, and during July (adults) and August (juveniles) on southward migration (Butler et al. 1987). We collected food abundance, safety, and site-use data from 17 mud- and sandflats in the Georgia Basin/Puget Sound (GB/PS) region in 2002 (Fig. 1). Three sites, Boundary Bay (BB), Roberts Bank (RB), and Skagit Bay (SB), were sampled in more than one migratory stage; all other sites were only sampled once during either the northward or southward migratory stage (Table 1).

Food Abundance

Western Sandpipers feed on a range of epibenthic and infaunal invertebrates (Wilson 1994, Elner and Seaman 2003, Mathot and Elner 2004). To measure food abundance for Western Sandpipers, we quantified benthic invertebrate abundance in sediment cores, according to the methods described by Sutherland et al. (2000). We collected samples within 500 m of the shoreline using a 60-mL syringe (26 mm inner diameter) to a depth of 40 mm. Samples were collected within 1.5 h of mud exposure by the receding tide and frozen within 1 h of collection. Cores were later thawed and rinsed through a 0.5-mm mesh sieve. Material on the sieve was preserved in vials with 85% ethanol. We identified and counted invertebrates in each vial to the lowest taxon possible under a 40X dissecting microscope. Each taxon sampled has been shown to be consumed by Western Sandpipers (Wolf 2001 and references therein). Wolf (2001) shows that the size range of invertebrates ingested by the Western Sandpiper are effectively sampled by this method, as the sizes of the amphipod *Corophium* recovered from sandpiper feces were within the size range of those in the core samples.

We collected and measured 15–146 cores at each site during at least one migratory stage (Table 1). Three sites were sampled in multiple migratory stages; invertebrate abundance did not differ among three migratory stages at Boundary Bay (ANOVA: $F_{2, 143} = 1.80, P = 0.17$) or among two stages at Roberts Bank (ANOVA: $F_{1, 52} = 2.89, P = 0.10$). At Skagit Bay, invertebrate abundance differed significantly between northward (mean invertebrate abundance \pm 95% confidence interval (C.I.) = $76.5 \pm 19.0, N = 30$) and southward adult (mean invertebrate abundance \pm 95% C.I. = $26.33 \pm 5.8, N = 30$) migratory stages ($F_{1, 58} = 25.4, P < 0.0001$).

Safety

Peregrine Falcons (*Falco peregrinus*), Merlins (*F. columbarius*), and other raptors hunt shorebirds, including Western Sandpipers (Page and Whitacre 1975, Whitfield 1985, Buchanan et al. 1988, Dekker 1988, Cresswell 1994, 1996), and achieve greatest hunting success using cover to conceal their approach (Whitfield 1985, Cresswell 1994, Whitfield 2003b, Dekker and Ydenberg 2004). As the shoreline poses a great deal of danger for foraging sandpipers, stopover sites that are small and have a large proportion of available foraging habitat close to cover are more dangerous than stopover sites that are large where sandpipers can spend large amounts of time feeding on the open mudflat far from the danger lurking along the shoreline (Pomeroy et al. 2006).

As an index of safety, we used the distance tool in ArcMap v.9.1 (ESRI 2005) to measure the furthest distance from the shoreline (as indicated by the upper water mark on marine charts (Nautical Data International 2005, National Oceanic and Atmospheric Administration (NOAA) 2005)) in the intertidal zone at each of the sites. For a large open mudflat or beach, this measure is the distance (m) between the shoreline (upper water mark) and the waterline (low water mark) at the widest point. For an enclosed bay surrounded by shoreline, the index of safety is measured as the distance from the shoreline to the midpoint of the widest distance across the bay.

Site Usage

Sites were classified as “used” or “unused” based on the intensity and frequency of use, using as sources published literature, unpublished reports, the Bird Studies Canada Coastal Waterbirds Survey (Bird Studies Canada, unpublished data, 2005) database, and local knowledge (Table 1). Sites were classified as “used” if hundreds or thousands of sandpipers on stopover are recorded there on a regular and annual basis (Butler et al. 1987, 1992, Buchanan 1988, Iverson et al. 1996, Evenson and Buchanan 1997, Warnock and Bishop 1998, Acevedo Seaman et al. 2006). Sites were considered “unused” if there are no records of Western Sandpipers there. A few sites (see Table 1) had single-day records up to 75 birds, but the sites are in fact visited only rarely by western sandpipers, and were therefore classified as “unused.”

Fig. 1. The Georgia Basin/Puget Sound region of British Columbia, Canada and Washington, USA.

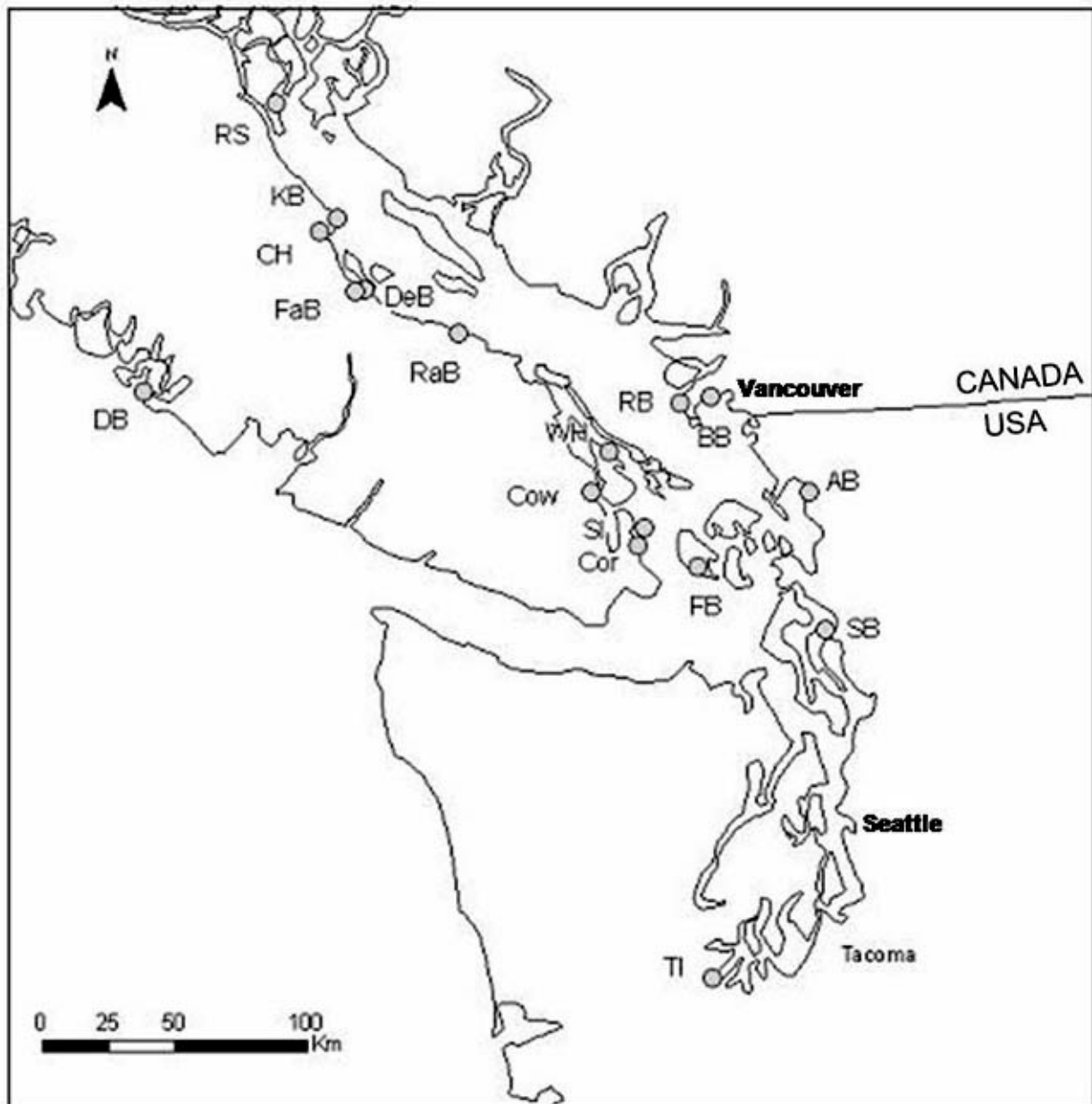


Table 1. The name and location of stopover sites sampled in the Georgia Basin/Puget Sound region, sample sizes for invertebrate cores and body-mass samples, and usage data for each migratory stage (N = northward, SA = southward adult, SJ = southward juvenile, *n* = number).

Site	Site Name	Location	Invertebrate core (<i>n</i>)	Body mass (<i>n</i>)	Usage	Peak Record	Migratory stage
AB	Alice Bay	Edison, WA	30	13	Used (Seaman 2003)	500	SA
BB	Boundary Bay	Delta, BC	146	N=63 SA=18 SJ=34	Used (Butler 1994)	>100 000	N, SA, SJ
CH	Comox Harbor	Comox, BC	15	N/A	Not used (BSC, unpublished data, 2005)	24	SA
Cor	Cordova Bay	Victoria, BC	15	N/A	Not used (BSC, unpublished data, 2005)	75	SA
Cow	Cowichan Bay	Duncan, BC	15	N/A	Not used	0	SA
DB	Doug Banks	Tofino, BC	25	28	Used (Butler et al. 1992)	5000	SJ
DeB	Deep Bay	Vancouver Island, BC	15	N/A	Not used (Dawe et al. 1998)	12	SA
FaB	Fanny Bay	Vancouver Island, BC	15	N/A	Not used (Dawe et al. 1998)	0	SA
FB	False Bay	San Juan Island, WA	25	7	Used (K. O'Reilly, unpublished data)	200	SJ
KB	Kye Bay	Comox, BC	15	N/A	Not used (BSC, unpublished data, 2005)	25	SA
RaB	Rath Trevor Beach	Parksville, BC	15	N/A	Not used (BSC, unpublished data, 2005)	20	SA
RB	Roberts Bank	Delta, BC	29	N=57 SA=18 SJ=66	Used (Butler 1994)	>100 000	N, SJ
RS	Rebecca Spit	Quadra Island, BC	15	N/A	Not used	0	SA
SB	Skagit Bay	Utsalady, WA	60	N=38, SA=25	Used (Evenson and Buchanan 1997)	10 000	N, SA
SI	Sidney Island	Sidney Island, BC	15	24	Used (Ydenberg et al. 2004)	1000	SJ
TI	Totten Inlet	Shelton, WA	31	27	Used (Evenson and Buchanan 1997)	2000	N
WH	Walker's Hook	Saltspring Island, BC	15	N/A	Not used (J. Booth, pers. comm.)	0	SA

At sites that were used during migration, we used mist nets to capture sandpipers to measure the body mass of individuals using the site (for sample sizes see Table 1). Western Sandpipers were removed from the mist nets immediately after capture and weighed within 10 min. Tarsae were measured using callipers (to the nearest 0.1 mm).

Statistical Analysis

We tested for differences between sites in food abundance and safety using analysis of variance (ANOVA). T-tests were used to test if food abundance and safety differed between used and unused sites. Separate logistic regressions were used to test whether food abundance, safety, or food abundance and safety affected stopover-site usage by sandpipers.

For analysis of body mass, we used analysis of covariance (ANCOVA) including tarsus length as a covariate to account for body size differences between individuals. As northbound migrants were significantly heavier (0.8 g) than southbound migrants ($P = 0.05$), we also included migratory stage (northward, southward adult, southward juvenile) in the analysis. Body-mass values at each site are reported as the least-squares mean. To investigate the relationship between site safety and state of individuals using a site, we then used these least-squares means of body mass at each site. Means and 95% C.I. are used throughout. JMP/IN V. 4.04 (SAS 2001) was used for all statistical analyses.

RESULTS

Of the 17 sites surveyed, eight were classified as “used” and nine as “unused” according to our criteria (Table 1). Food abundance and safety measures both varied widely between study sites. Food abundance ranged from means of eight to 204 invertebrates core⁻¹ (ANOVA: $F_{16, 479} = 16.3$, $P < 0.0001$). Invertebrate densities were on average three times greater at used than at unused sites (mean invertebrate density/core \pm 95% C.I.; used sites = 99.4 ± 32.2 ; unused sites = 31.9 ± 30.4 , ANOVA: $F_{1, 15} = 9.26$, $P = 0.008$). The relative abundance of invertebrate taxa sampled at each site is provided in Appendix 1. The safety index ranged from 75 m to 4560 m, with used sites on average three times safer (mean distance from shore \pm 95% C.I. = 1932.5

± 845.9 m) than unused sites (mean distance from shore \pm 95% C.I. = 581.7 ± 797.5 m, ANOVA: $F_{1, 15} = 5.4$, $P = 0.03$, $N = 17$).

In a logistic regression model including food as the only independent variable, usage of 12 of the 17 sites was predicted correctly (Table 2. $\chi^2 = 8.51$, d.f. = 1, $R(U)^2 = 0.36$, $P = 0.0035$). The logistic regression model including safety as the only independent variable also correctly predicted usage of 12 of 17 sites ($\chi^2 = 5.78$, d.f. = 1, $R(U)^2 = 0.25$, $P = 0.02$), although the identity of the 12 sites differed (Table 2). The logistic regression including both food abundance and safety as explanatory variables performed better than either of these models, correctly predicting the usage of 14 of 17 sites (Fig. 2. $\chi^2 = 10.24$, d.f. = 2, $R(U)^2 = 0.44$, $P = 0.006$).

The body mass of individuals (controlled for structural size and migratory stage) was greater at stopover sites with greater safety (Fig. 3. Nonlinear Regression: LS mean body mass = $26.6 * (1 - e^{-0.0061 * \text{safetyindex}})$, $R^2 = 0.55$, $P = 0.03$), but was not significantly related to food abundance ($P = 0.41$).

DISCUSSION

Our analysis shows that including both food abundance and safety as factors improved our ability to predict which stopover sites were used by migrating sandpipers and which were not. Either factor on its own was also able to predict stopover-site usage, but migrant sandpipers used sites that were dangerous only if food abundance there was high, and used sites with low food abundance only if they were very safe. To our knowledge, no other study has investigated the usage of stopover sites by migrants or compared multiple site attributes at both used and unused sites on a regional scale. Our results also confirm that, on this scale, the usage of stopover sites depends on the fuel load of individual sandpipers, previously shown in a comparison of just two sites by Ydenberg et al. (2002). Sandpipers on dangerous stopover sites carry on average less fuel than those on safer sites. We hypothesize that this occurs because sandpipers with adequate fuel loads do not need to risk feeding in dangerous places.

Our best model incorrectly predicted usage of three of the 17 sites, and an examination of their

Table 2. Actual usage at sample sites on migration by sandpipers, and usage as predicted by each logistic regression model. (Y = yes; N = no.)

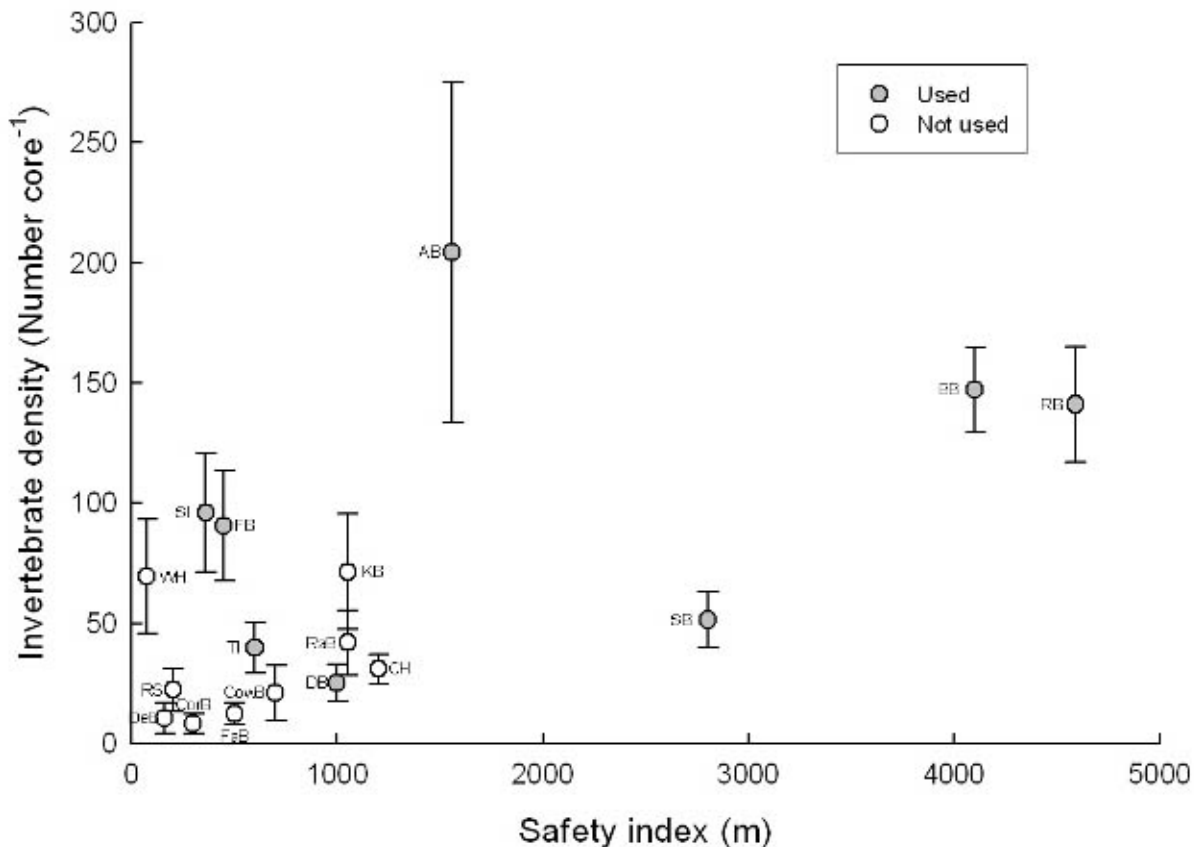
Site	Used	Predicted used by food only	Predicted used by danger only	Predicted used by food and danger
AB	Y	Y	Y	Y
BB	Y	Y	Y	Y
CH	N	N	Y	N
CorB	N	N	N	N
CowB	N	N	N	N
DB	Y	N	N	N
DeB	N	N	N	N
FaB	N	N	N	N
FB	Y	Y	N	Y
KB	N	Y	N	Y
RaB	N	N	N	N
RB	Y	Y	Y	Y
RS	N	N	N	N
SB	Y	N	Y	Y
SI	Y	Y	N	Y
TI	Y	N	N	N
WH	N	Y	N	N

characteristics can help to understand these assignment errors. Two sites (Doug Banks and Totten Inlet) were classified by the model as “unused,” but are in fact used regularly by sandpipers on stopover. The cover surrounding these two sites differs from that lining the shoreline of the other six “used” sites, which consists of low marsh grasses whereas Doug Banks and Totten Inlet are surrounded by tall coniferous trees. From the perspective of a migrant sandpiper, these sites may be functionally safe, as falcons do not appear to use these trees for attack cover, but hunt by attacking over the open mudflat, much like the attack strategy

they employ at larger sites (J. Buchanan, pers. comm.). This difference suggests that cover type might be an important factor influencing stopover-site usage by migrants and may be used, in addition to distance from cover, to remotely assess the danger of a potential site.

The third incorrectly classified site was Kye Bay, which, according to the model, should be used by sandpipers due to its high food abundance; however, there is no record of usage by Western Sandpipers. A noticeable difference between this and all other used stopover sites is the relatively high percentage

Fig. 2. The mean invertebrate density (\pm 95% C.I.) and safety index for sites that are used and not used by Western Sandpipers on migration.



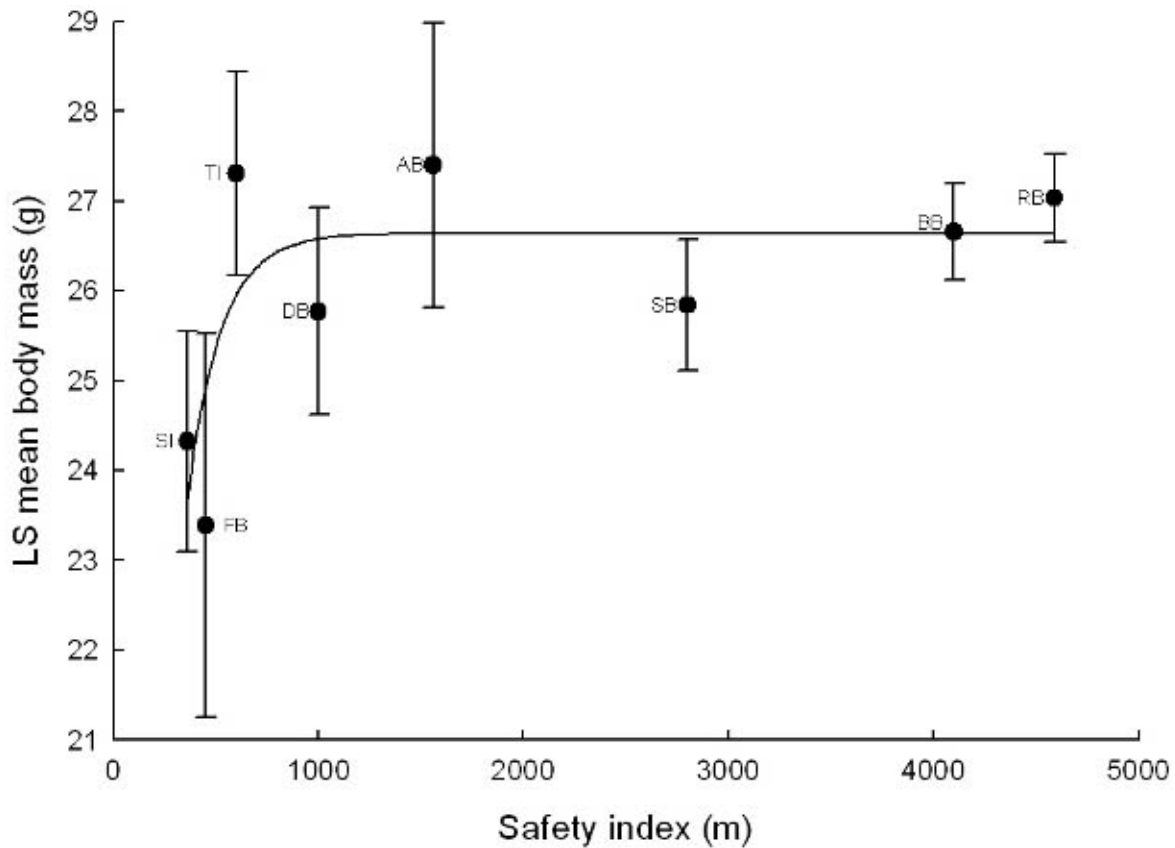
(60%) of nematodes (Phylum Nematoda, Class Adenophorea) among the potential prey items, compared with a maximum of 46% and an average of 25% nematodes among used sites (Appendix 1). Although all prey taxa presented here have been documented as Western Sandpiper prey (Wolf 2001), there is evidence that nematodes are less preferred by Western Sandpipers than other prey items such as polychaetes (Sutherland et al. 2000). It is possible that prey type may also be an important factor influencing stopover-site selection by migrants and may be used, in addition to prey abundance, to remotely assess the food availability of a potential site.

One possible source of bias in this study is from our measure of food at each site. We measured

invertebrate abundance from core samples extracted within 500 m of the shoreline at each site; however, this methodology might not accurately represent food abundance across entire sites. At Boundary Bay, for example, there is a strong declining gradient in food abundance as distance from the shoreline increases (Pomeroy 2006). For large sites where these gradients are likely to occur (Swennen et al. 1982, Reise 1985, Kennish 1986), we might be overestimating food abundances, whereas our measures of food abundance at smaller sites are likely more accurate. Despite this, our analyses remain predictive.

This study indicates that both food and danger attributes of a site influence stopover-site usage by migrants. Stopover-site selection by migrants is

Fig. 3. The relationship between the safety index measured at stopover sites and the least-squares mean body mass (\pm 95% C.I.) of sandpipers captured. The line is that predicted from the nonlinear regression (LS mean body mass = $26.6 * (1 - e^{-0.0061 * \text{safety index}})$)).



thought to be hierarchical (Hutto 1985). To avoid the costs of sampling potential stopover sites, migrants likely use features of the environment to approximate levels of food abundance and predation danger at the site level to decide whether or not to use the site. For example, high densities of invertebrates and biofilm, on which sandpipers feed (Kuwaie et al. 2008), are often associated with fine-grained sediments (Kennish 1990, Yates et al. 1993). It is possible that migratory shorebirds can visually assess food abundance of potential stopover sites from the air, e.g., by the reflectance or sediment properties of a beach or mudflat (Rainey et al. 2003, Pomeroy and Butler 2005). To remotely assess predation danger at a site, migrants may use proximity to cover as a cue. Large mudflats with a vast expanse of foraging habitats distant from the

shoreline are indicative of a relatively safe site for a migrant shorebird, whereas a small mudflat surrounded by marsh grass would be deemed a potentially dangerous place to stop. Upon selecting a stopover site, migrants can more accurately assess the levels of food abundance, and gauge the level of danger there based on their encounters with predators, and they can make appropriate adjustments to anti-predator behaviors to carefully balance food and danger over short temporal and small spatial scales. This work suggests that migrants assess food and danger attributes of stopover sites remotely and that these attributes affect which sites they will use.

Assessment of the conservation value of stopover sites for migrants should evaluate measures of both

food and danger. As migration is energetically expensive, stopover-site quality is often associated with the quantity or quality of food resources (Yong et al. 1998, Baker et al. 2004, Placyk and Harrington 2004, Battley et al. 2005, Stillman et al. 2005a, van Gils et al. 2005, Smith et al. 2007). However, migration is also dangerous. As danger from predators increases, migrant birds have been shown to alter their behavior at stopover sites by decreasing length of stay at the site (Ydenberg et al. 2004), allocating more feeding time to vigilance (Cimprich et al. 2005, Pomeroy 2006), and carrying lower fuel loads (Ydenberg et al. 2002, 2004, Schmaljohann and Dierschke 2005). Predation danger influences habitat usage by migrants within and between stopover sites. Assessments of stopover-site quality then must include both food and danger attributes of sites. Furthermore, stopover-site quality is condition dependent. Whereas heavy individuals with ample fuel reserves may prefer safe sites with high bird densities despite few food opportunities, lean birds would be more likely to consider a high-quality site as one where it could fatten quickly, despite the added danger. Our results suggest that a variety of site types (i.e., both high food and high safety) need to be conserved.

According to this study, usage of stopover sites by migrants will change if levels of food and/or danger change depending on a) the magnitude of the change, and b) the level of the other attribute. For example, if food abundance declines at a safe site, usage by migrants might not change, whereas if the site is dangerous it might no longer be used. At the landscape scale, increases in predator abundance would likely cause migrants to shift usage from small, dangerous sites to larger, safer ones (e.g., Ydenberg et al. 2004, Taylor et al. 2007). Moreover, changes in stopover-site usage will be condition dependent (Ydenberg et al. 2002, 2004). A decrease in food abundance at a dangerous site will have a greater effect on lean birds, whereas increasing danger at safe sites will have a greater effect on heavy birds. If we neglect the role of predators and the effect they have on stopover-site usage, the behavioral adaptations used by migrants to avoid mortality by predation, including increasing vigilance, decreasing length of stay, and adjusting habitat usage, could instead be attributed to declines in food abundance at a site or to population decline (Ydenberg et al. 2004). Efforts to conserve habitat for migrants must consider the role of food abundance and predation danger in stopover-site usage.

Stopover-site conservation is essential to maintain populations of migratory birds. The importance of identifying, prioritizing, and protecting stopover habitat for birds is gaining increasing recognition (Myers et al. 1987, Donovan et al. 2002, Mehlman et al. 2005, Skagen 2006). Many bird conservation programs identify and prioritize sites of importance based on the number of individuals (or proportion of a population) using a particular site (e.g., Western Hemisphere Shorebird Reserve Network (WHSRN), Important Bird Area (IBA)). However, this approach may overemphasize the importance of large (safe) sites that can support significant numbers of birds. Mehlman et al. (2005) assert that efforts to conserve stopover habitat for landbirds should target a mosaic of different types of stopover habitats. The authors suggest three categories of stopover sites representing points on a continuum of their capacity to meet the needs of migrants. These categories range from fire escapes (infrequently used with few available resources, but vital in emergencies) to convenience stores (mid-quality stopover sites that are used for rest, and replenishment of some fuel, but where food resources might be low and/or predation danger high), to full-service hotels (high-quality sites with abundant resources where migrants can safely rest and refuel for the next leg of migration). In our study system, the sites with ample food and safety would be considered full-service hotels, and sites that are high in either food or safety could be thought of as convenience stores, whereas sites that are low in food and safety may be potential fire escapes. Within each of these categories, the sites can then easily be ranked based on their food and safety attributes, and prioritized accordingly for conservation.

We show here evidence that migrant Western Sandpipers select stopover sites according to trade-offs between food abundance and predation danger. Furthermore, usage of these sites depends on the state of an individual. This study suggests that migrants use habitat features, such as cover, to assess predation danger, and that they mediate their probability of mortality from predation by adjusting habitat usage on a landscape scale. Results such as these can be applied to predict the behavior of migrants at stopover sites, and usage of stopovers if food and/or danger attributes at a site change. Furthermore, stopover-site usage by migrants depends on state. As danger from predators changes on the landscape, the state of the birds that use those sites might also change (Ydenberg et al. 2004). This

study indicates that to identify migration stopover sites for conservation, both food and danger attributes must be considered.

Responses to this article can be read online at:
<http://www.ace-eco.org/vol3/iss1/art7/responses/>

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LITERATURE CITED

- Acevedo Seaman, D. A., C. G. Guglielmo, R. W. Elner, and T. D. Williams.** 2006. Landscape-scale physiology: site differences in refueling rates indicated by plasma metabolite analysis in free-living, migratory sandpipers. *Auk* **123**:563–574.
- Baker, A. J., P. M. González, T. Piersma, L. J. Niles, I. d. L. S. d. Nascimento, P. W. Atkinson, N. A. Clark, C. D. T. Minton, M. K. Peck, and G. Aarts.** 2004. Rapid population decline in red knots: fitness consequences of decreased refuelling rates and late arrival in Delaware Bay. *Proceedings of the Royal Society of London: Series B* **271**:875–882.
- Battley, P. F., D. I. Rogers, J. A. van Gils, T. Piersma, C. J. Hassell, A. Boyle, and Y. Hong-Yan.** 2005. How do red knots *Calidris canutus* leave Northwest Australia in May and reach the breeding grounds in June? Predictions of stopover times, fuelling rates and prey quality in the Yellow Sea. *Journal of Avian Biology* **36**:494–500.
- Buchanan, J. B.** 1988. The abundance and migration of shorebirds at two Puget Sound estuaries. *Western Birds* **19**:67–78.
- Buchanan, J. B., C. T. Schick, L. A. Brennan, and G. Herman.** 1988. Merlin predation on wintering dunlins; hunting success and dunlin escape tactics. *Wilson Bulletin* **100**:108–118.
- Burke, D. M., and E. Nol.** 1998. Influence of food abundance, nest-site habitat, and forest fragmentation on breeding ovenbirds. *Auk* **115**:96–104.
- Butler, R. W.** 1994. Distribution and abundance of Western Sandpipers, Dunlins, and Black-bellied Plovers in the Fraser River estuary. Pages 18–23 in R. W. Butler and K. Vermeer, editors. *Abundance and distribution of birds in estuaries in the Strait of Georgia*. Occasional Paper Number 83, Canadian Wildlife Service, Ottawa, Ontario, Canada.
- Butler, R. W., A. Dorst, and M. A. Hobson.** 1992. Seasonal abundance and biomass of birds in eelgrass habitats in Browning Passage on the west coast of Vancouver Island. Pages 109–113 in K. Vermeer, R. W. Butler, and K. H. Morgan, editors. *The ecology and status of marine and shoreline birds on the west coast of Vancouver Island, British Columbia*. Occasional Paper No. 75, Canadian Wildlife Service, Ottawa, Ontario, Canada.
- Butler, R. W., G. W. Kaiser, and G. E. J. Smith.** 1987. Migration chronology, length of stay, sex ratio, and weight of western sandpipers (*Calidris mauri*) on the south coast of British Columbia. *Journal of Field Ornithology* **58**:103–111.
- Cimprich, D. A., M. S. Woodrey, and F. R. Moore.** 2005. Passerine migrants respond to variation in predation risk during stopover. *Animal Behaviour* **69**:1173–1179.
- Clark, C. W. and R. W. Butler.** 1999. Fitness components of avian migration: a dynamic model of Western Sandpiper migration. *Evolutionary Ecology Research* **1**:443–457.

- Colwell, M. A. and S. L. Landrum.** 1993. Nonrandom shorebird distribution and fine-scale variation in prey abundance. *Condor* **95**:94–103.
- Cresswell, W.** 1994. Age-dependent choice of redshank (*Tringa totanus*) feeding location: profitability or risk? *Journal of Animal Ecology* **63**:589–600.
- Cresswell, W.** 2008. Non-lethal effects of predation in birds. *Ibis* **150**:3–17.
- Cresswell, W.** 1996. Surprise as a winter hunting strategy in Sparrowhawks *Accipiter nisus*, Peregrines *Falco peregrinus* and Merlins *F. columbarius*. *Ibis* **138**:684–692.
- Dawe, N. K., R. Buechert, and D. E. C. Trethewey.** 1998. Bird use of Baynes Sound—Comox Harbour, Vancouver Island 1980–1981. Technical Report Number 286, Canadian Wildlife Service, Pacific and Yukon Region, Delta, British Columbia, Canada.
- Dekker, D.** 1988. Peregrine falcon and merlin predation on small shorebirds and passerines in Alberta. *Canadian Journal of Zoology* **66**:925–928.
- Dekker, D., and R. C. Ydenberg.** 2004. Raptor predation on wintering Dunlins in relation to the tidal cycle. *Condor* **106**:415–419.
- Donovan, T. M., C. J. Beardmore, D. N. Bonter, J. D. Brawn, R. J. Cooper, J. A. Fitzgerald, R. Ford, S. A. Gauthreaux, T. L. George, W. C. Hunter, T. E. Martin, J. Price, K. V. Rosenberg, P. D. Vickery, and T. B. Wigley.** 2002. Priority research needs for the conservation of Neotropical migrant landbirds. *Journal of Field Ornithology* **73**:329–339.
- Elnor, R. W., and D. A. Seaman.** 2003. Calidrid conservation: unrequited needs. *Wader Study Group Bulletin* **100**:30–34.
- ESRI.** 2005. ArcMap V. 9.1. Environmental Systems Research Institute, Inc., Toronto, Ontario, Canada.
- Evenson, J. R., and J. B. Buchanan.** 1997. Seasonal abundance of shorebirds at Puget Sound estuaries. *Washington Birds* **6**:34–62.
- Gill, J. A., K. Norris, P. M. Potts, T. G. Gunnarsson, P. W. Atkinson, and W. J. Sutherland.** 2001. The buffer effect and large-scale population regulation in migratory birds. *Nature* **412**:436–438.
- Hicklin, P. W., and P. C. Smith.** 1984. Selection of foraging sites and invertebrate prey by migrant Semipalmated Sandpipers, *Calidris pusilla* in Minas Basin, Bay of Fundy. *Canadian Journal of Zoology* **62**:2201–2210.
- Hutto, R. L.** 1985. Habitat selection by nonbreeding, migratory land birds. Pages 455–476 in M. L. Cody, editor. *Habitat selection in birds*. Academic Press, New York, New York, USA.
- Iverson, G. C., S. E. Warnock, R. W. Butler, M. A. Bishop, and N. Warnock.** 1996. Spring migration of Western Sandpipers along the Pacific coast of North America: a telemetry study. *Condor* **98**:10–21.
- Kennish, M. J.** 1986. *Ecology of estuaries: volume I Physical and chemical aspects*. CRC Press, Boca Raton, Florida, USA.
- Kennish, M. J.** 1990. *Ecology of estuaries: volume II Biological aspects*. CRC Press, Boca Raton, Florida, USA.
- Kuwa, T., P. G. Beninger, P. Decottignies, K. J. Mathot, D. R. Lund, and R. W. Elnor.** 2008. Biofilm grazing in a higher vertebrate: the Western Sandpiper *Calidris mauri*. *Ecology* **89**:599–606.
- Lank, D. B., R. W. Butler, J. Ireland, and R. C. Ydenberg.** 2003. Effects of predation danger on migration strategies of sandpipers. *Oikos* **103**:303–319.
- Latta, S. C., and J. Faaborg.** 2002. Demographic and population responses of cape may warblers wintering in multiple habitats. *Ecology* **83**:2502–2515.
- Loria, D. E., and F. R. Moore.** 1990. Energy demands of migration on red-eyed vireos, *Vireo olivaceus*. *Behavioral Ecology* **1**:24–35.
- Mathot, K. J., and R. W. Elnor.** 2004. Evidence for sexual partitioning of foraging mode in Western Sandpipers (*Calidris mauri*) during migration. *Canadian Journal of Zoology* **82**:1035–1042.

- Mehlman, D. W., S. E. Mabey, D. N. Ewert, C. Duncan, B. Able, D. Cimprich, R. D. Sutter, and M. Woodrey.** 2005. Conserving stopover sites for forest-dwelling migratory land birds. *Auk* **122**:1281–1290.
- Moore, F. R.** 1994. Resumption of feeding under risk of predation: effect of migratory condition. *Animal Behaviour* **48**:975–977.
- Moore, F. R., S. A. Gauthreaux, Jr., P. Kerlinger, and T. R. Simons.** 1995. Habitat requirements during migration: Important link in conservation. Pages 121–144 in T. E. Martin and D. M. Finch, editors. *Ecology and management of neotropical migratory birds: a synthesis and review of critical issues*. Oxford University Press, New York, New York, USA.
- Myers, J.P., R. I. G. Morrison, P. Z. Antas, B. A. Harrington, T. E. Lovejoy, M. Sallaberry, S. E. Senner, and A. Tarak.** 1987. Conservation strategy for migratory species. *American Scientist* **75**:19–26.
- National Oceanic and Atmospheric Administration (NOAA).** 2005. *NOAA Navigational Charts*. Office of Coast Survey, National Ocean Survey, Silver Spring, Maryland, USA.
- Nautical Data International Inc.** 2005. *Canadian Hydrographic Service Nautical Charts*. Fisheries and Oceans Canada, Ottawa, Ontario, Canada.
- Page, G., and D. F. Whitacre.** 1975. Raptor predation on wintering shorebirds. *Condor* **77**:73–83.
- Piersma, T., Y. Verkuil, and I. Tulp.** 1994. Resources for long-distance migration of knots *Calidris canutus islandica* and *C. c. canutus*: how broad is the temporal exploitation window of benthic prey in the western and eastern Wadden Sea? *Oikos* **71**:393–407.
- Placyk, J. S. Jr., and B. A. Harrington.** 2004. Prey abundance and habitat use by migratory shorebirds at coastal stopover sites in Connecticut. *Journal of Field Ornithology* **75**:223–231.
- Placyk, J. S., and B. A. Harrington.** 2004. Prey abundance and habitat use by migratory shorebirds at coastal stopover sites in Connecticut. *Journal of Field Ornithology* **75**:223–231.
- Pomeroy, A. C.** 2006. Tradeoffs between food abundance and predation danger in spatial usage of a stopover site by western sandpipers, *Calidris mauri*. *Oikos* **112**:629–637.
- Pomeroy, A. C., and R. W. Butler.** 2005. Color infrared photography is not a good predictor of macro invertebrate abundance on mudflats used by shorebirds. *Waterbirds* **28**:1–7.
- Pomeroy, A. C., R. W. Butler, and R. C. Ydenberg.** 2006. Experimental evidence that migrants adjust usage at a stopover site to trade off food and danger. *Behavioral Ecology* **17**:1041–1045.
- Rainey, M. P., A. N. Tyler, D. J. Gilvear, R. G. Bryant, and P. McDonald.** 2003. Mapping intertidal estuarine sediment grain size distributions through airborne remote sensing. *Remote Sensing of the Environment* **84**:480–490.
- Reise, K.** 1985. *Tidal flat ecology: an experimental approach to species interactions*. Springer-Verlag, Berlin, Germany.
- Robinson, S. K., F. R. Thompson III, T. M. Donovan, D. R. Whitehead, and J. Faaborg.** 1995. Regional forest fragmentation and the nesting success of migratory birds. *Science* **267**:1987–1990.
- Rodewald, A. D., and R. H. Yahner.** 2001. Avian nest success in forested landscapes: influence of landscape composition, stand and nest-patch microhabitat, and biotic interactions. *Auk* **118**:1018–1028.
- Sapir, N., Z. Abramsky, E. Shochat, and I. Izhaki.** 2004. Scale-dependent habitat selection in migratory frugivorous passerines. *Naturwissenschaften* **91**:544–547.
- SAS.** 2001. JMPIN V. 4.0.4. SAS Institute, Cary, North Carolina, USA.
- Schmaljohann, H., and V. Dierschke.** 2005. Optimal bird migration and predation risk: a field experiment with northern wheatears (*Oenanthe oenanthe*). *Journal of Animal Ecology* **74**:131–138.
- Schneider, D. C., and B. A. Harrington.** 1981. Timing of shorebird migration in relation to prey depletion. *Auk* **98**:801–811.

- Seaman, D. A.** 2003. Landscape physiology: plasma metabolites, fattening rates and habitat quality in migratory Western Sandpipers. Thesis, Biological Sciences Department, Simon Fraser University, Burnaby, British Columbia, Canada.
- Sillett, T. S., and R. T. Holmes.** 2002. Variation in survivorship of a migratory songbird throughout its annual cycle. *Journal of Animal Ecology* **71**:296–308.
- Skagen, S. K.** 2006. Migration stopovers and the conservation of arctic-breeding calidridine sandpipers. *Auk* **123**:313–322.
- Skagen, S. K., and F. J. Knopf.** 1993. Toward conservation of midcontinental shorebird migrations. *Conservation Biology* **7**:533–541.
- Smith, R. J., F. R. Moore, and C. A. May.** 2007. Stopover habitat along the shoreline of northern Lake Huron, Michigan: emergent aquatic insects as a food resource for spring migrating landbirds. *Auk* **124**:107–121.
- Stillman, R. A., A. D. West, J. D. Goss-Custard, S. McGrorty, N. J. Frost, D. J. Morrissey, A. J. Kenny, and A. L. Drewitt.** 2005a. Predicting site quality for shorebird communities: a case study on the Humber estuary, UK. *Marine Ecology Progress Series* **305**:203–217.
- Stillman, R. A., A. D. West, J. D. Goss-Custard, S. McGrorty, N. J. Frost, D. J. Morrissey, A. J. Kenny, and A. L. Drewitt.** 2005b. Predicting site quality for shorebird communities: a case study on the Humber estuary, UK. *Marine Ecology Progress Series* **305**:203–217.
- Sutherland, T. F., P. C. F. hepherd, and R. W. Elner.** 2000. Predation on meiofaunal and macrofaunal invertebrates by Western Sandpipers (*Calidris mauri*): evidence from dual foraging modes. *Marine Biology* **137**:983–993.
- Swennen, C., P. Duiven, and A. L. Spaans.** 1982. Numerical density and biomass of macrobenthic animals living in the intertidal zone of Surinam, South America. *Netherlands Journal of Zoology* **15**:406–418.
- Taylor, C. M., D. B. Lank, A. C. Pomeroy, and R. C. Ydenberg.** 2007. Relationship between stopover site choice of migrating sandpipers, their population status, and environmental stressors. *Israel Journal of Ecology and Evolution* **53**: 245–261.
- van Gils, J. A., P. F. Battley, T. Piersma, and R. Drent.** 2005. Reinterpretation of gizzard sizes of red knots world-wide emphasises overriding importance of prey quality at migratory stopover sites. *Proceedings of the Royal Society of London: Series B* **272**:2609–2618.
- Warnock, N., and M. A. Bishop.** 1998. Spring stopover ecology of migrant western sandpipers. *Condor* **100**:456–467.
- Weber, T. P., B. J. Ens, and A. I. Houston.** 1998. Optimal avian migration: a dynamic model of fuel stores and site use. *Evolutionary Ecology* **12**:377–401.
- Whitfield, D. P.** 1985. Raptor predation on wintering waders in southeast Scotland. *Ibis* **127**:552–558.
- Whitfield, D. P.** 2003a. Predation by Eurasian sparrowhawks produces density-dependent mortality of wintering redshanks. *Journal of Animal Ecology* **72**:27–35.
- Whitfield, D. P.** 2003b. Redshank *Tringa totanus* flocking behaviour, distance from cover and vulnerability to sparrowhawk *Accipiter nisus* predation. *Journal of Avian Biology* **34**:163–169.
- Wilson, W. H. J.** 1994. The effects of episodic predation by migratory shorebirds in Grays Harbor, Washington. *Journal of Experimental Marine Biology and Ecology* **177**:15–25.
- Wolf, N.** 2001. Foraging ecology and stopover site selection of migrating Western Sandpipers (*Calidris mauri*). Thesis, Biological Sciences Department, Simon Fraser University, Burnaby, British Columbia, Canada.
- Yates, M. G., J. D. Goss-Custard, S. McGrorty, K. H. Lakhani, S. E. A. L. Durell, R. T. Clarke, W. E. Rispin, I. Moy, T. Yates, R. A. Plant, and A. J. Frost.** 1993. Sediment characteristics, invertebrate densities and shorebird densities on the inner banks of the Wash. *Journal of Applied Ecology* **30**:599–614.
- Ydenberg, R. C., R. W. Butler, D. B. Lank, C. G. Guglielmo, M. Lemon, and N. Wolf.** 2002. Trade-

offs, condition dependence and stopover site selection by migrating sandpipers. *Journal of Avian Biology* **33**:47–55.

Ydenberg, R. C., R. W. Butler, D. B. Lank, B. D. Smith, and J. Ireland. 2004. Western sandpipers have altered migration tactics as peregrine falcon populations have recovered. *Proceedings of the Royal Society of London: Series B* **271**:1263–1269.

Yong, W., D. M. Finch, F. R. Moore, and J. F. Kelly. 1998. Stopover ecology and habitat use of migratory Wilson's warblers. *Auk* **115**:829–842.

Zanette, L., P. Doyle, and S. M. Trémont. 2000. Food shortage in small fragments: evidence from an area-sensitive passerine. *Ecology* **81**:1654–1666.

Appendix 1. The percentage of each taxon represented at each site.

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